COMMUNITY ECOLOGY - ORIGINAL RESEARCH

# From selection to complementarity: the shift along the abiotic stress gradient in a controlled biodiversity experiment

Jiang Wang • Chong B. Zhang • Tong Chen • Wei H. Li

Received: 27 August 2011 / Accepted: 12 June 2012 / Published online: 13 July 2012 © Springer-Verlag 2012

Abstract The effects of principal mechanisms (selection and complementarity) of biodiversity on ecosystem functionality have been well studied. However, it remains unknown how environmental conditions affect the relative strength of these two mechanisms. To answer this question, a controlled pot experiment was conducted in which species diversity was manipulated in low (natural soil) and high stress (mine tailing) plots, respectively. Our results demonstrate that the principal mechanism underlying the increasing biomass shifts from the selection to complementarity with increasing abiotic stress. The shift occurs because species interactions varied with increasing abiotic stress. Competition prevails in low stress plots, while facilitation dominates in high stress plots. In low stress plots, the monoculture biomass of a specific species is a good indicator of the competitive ability of that species in the mixture, and the dominant species significantly affects the plot biomass. In high stress plots, the tolerance indexes of all individual species increase with the manipulated

Communicated by Scott Collins.

Electronic supplementary material The online version of this article (doi:[10.1007/s00442-012-2400-2\)](http://dx.doi.org/10.1007/s00442-012-2400-2) contains supplementary material, which is available to authorized users.

J. Wang - C. B. Zhang - T. Chen School of Life Science, Taizhou University, Linhai, People's Republic of China

W. H. Li  $(\boxtimes)$ 

Key Laboratory of Ecology and Environmental Science, Guangdong Higher Education, School of Life Sciences, South China Normal University, Guangzhou, People's Republic of China e-mail: 541zf@163.com

species richness, providing clear evidence for the increasing role of facilitation.

Keywords Selection effect · Complementarity effect · Species interaction - Abiotic stress - Biomass

# Introduction

Laboratory and field experiments have shown that plant diversity positively affects productivity (Wardle [1999](#page-8-0); Schwartz et al. [2000;](#page-8-0) Špaèková and Lepš<sup>[2001](#page-8-0)</sup>; Hector [2002](#page-7-0); Hector et al. [2002](#page-7-0); Pfisterer et al. [2004;](#page-8-0) Hooper et al. [2005](#page-7-0); Spehn et al. [2005](#page-8-0); Fargione et al. [2007\)](#page-7-0). This can be explained by two major hypotheses: the selection effect (the greater probability of including a species or combinations of species with the trait of higher than average biomass in a randomly assembled high diversity community) and the complementarity effect (resource partitioning or facilitative interaction) (Huston [1997](#page-8-0); Loreau and Hector [2001](#page-8-0); Huston and McBride [2002](#page-8-0)). The selection and complementarity effects are not inherently exclusive, and both effects can operate simultaneously to affect produc-tivity (Spaèková and Lepš<sup>[2001](#page-8-0)</sup>). Consequently, which principal mechanism causes the positive effects of biodiversity on productivity has been actively debated (Huston et al. [2000;](#page-8-0) Kaiser [2000;](#page-8-0) Naeem [2000;](#page-8-0) Wardle et al. [2000](#page-8-0); Fargione et al. [2007\)](#page-7-0).

The debate highlights the need to separate and quantify the selection and complementarity effects. Loreau ([1998\)](#page-8-0) showed that over-yielding (i.e. higher productivity of a mixture when compared with the monoculture productivity of the most productive species) cannot be accounted for by the selection effect. Hence, over-yielding can be considered as evidence of additional mechanisms beyond the

selection effect. Loreau and Hector ([2001\)](#page-8-0) devised a statistical method to differentiate the selection and complementarity effects through comparing the mixture yield with the expected yield based on monocultures of component species. This statistical technique and the comparison of species performances of mixtures with monocultures have become important tools in assessing the effects of biodiversity change (Hector et al. [2002\)](#page-7-0). However, previous results indicated that the relationship between species production in monocultures and that in the mixtures was more complex (Hooper and Vitousek [1998](#page-7-0); Dukes [2001](#page-7-0); Engelhardt and Ritchie [2001](#page-7-0); Hector et al. [2002](#page-7-0)). The complex relationships may be attributed to the change in species interactions as a result of a trade-off between resource allocation to growth and to competition, a result of different competitive abilities to different resources of a specific species, or the consequence of the variation in a given environment in which resources are being competed for (Tilman [1988](#page-8-0)). To explain the change in selection and complementarity effects, a mechanistic understanding of species interactions is needed.

The stress gradient hypothesis predicts that the relative importance of facilitation and competition varies inversely across the abiotic stress gradient, with the facilitation being the dominant species interaction under high stress conditions (Bertness and Callaway [1994](#page-7-0)). However, a critical reappraisal of the hypothesis is needed because some experimental studies refuted its predictions (Liancourt et al. [2005](#page-8-0); Maestre et al. [2005\)](#page-8-0). These contradictory results may be related to the tolerance levels of species to certain abiotic stress. If stress-tolerant species are the experimental targets, the abiotic stress will not be particularly harsh for these species. As a result, competition may be expected. Consequently, species' interactions may vary from competition to facilitation as a function of the tolerance of a species to particular abiotic conditions (Bertness et al. [1992](#page-7-0); Hacker and Bertness [1999](#page-7-0)). Although the selection and complementarity effects are potentially relevant, they have different implications and are applied to different environmental conditions (Loreau [2000](#page-8-0); Loreau and Hector [2001](#page-8-0)). The selection effect may be particularly important in environments where inter-specific competition is promoted, and hence one or a few species have strong individual control over productivity (Fridley [2001\)](#page-7-0). The complementarity effect will be dominant in environments where the facilitation is the dominant interaction (Chu et al. [2008](#page-7-0)). However, relatively few studies have empirically examined how these interactions affect the functioning of whole ecosystems (Mulder et al. [2001;](#page-8-0) Kikvidze et al. [2005](#page-8-0)). Therefore, the importance of facilitation and competition for maintaining ecosystem functioning is largely unknown (Callaway [2007\)](#page-7-0).

Mine tailings are mechanically, physically, chemically and biologically deficient for plant species (Vega et al. [2006](#page-8-0); Wang et al. [2011\)](#page-8-0), and are characterized by instability and limited cohesion, a low content of nutrients and organic matter, and high levels of heavy metals (He et al. [2005](#page-7-0); Wang et al. [2011](#page-8-0)). To assess the relative importance of selection and complementarity effects across gradients of abiotic stress, we performed a controlled pot experiment in which species diversity was manipulated in low (natural soil) and high (mine tailing) stress environments, respectively. Specifically, we aimed to determine: (1) whether the relative importance of selection and complementarity effects varied with increasing abiotic stress; (2) whether the shift in the strength of both the selection and complementarity effects was due to the change in species interaction.

## Materials and methods

A plant species survey was conducted at Huangyan Pb/Zn mine tailing (28°34'23"N, 120°53'44"E) on 14 April 2008 (Online Resource 1). Eight species (Bidens pilosa Linn., Phytolacca americana Linn., Commelina communis Linn., Mirabilis jalapa Linn., Chenopodium ambrosioides Linn., Solanum nigrum Linn., Brassica campestris Linn. and Xanthium sibiricum Patrin ex Widder) were randomly selected for this experiment. P. americana is a perennial grass, and the other seven species are annual grasses. Seeds of the eight grass species were collected from the plants growing on the mine tailings or the surrounding area. The seeds of these species were sown in trays on 5–7 April 2009, and seedlings were transplanted 2 months after germination. Tailing soils was randomly collected from the bare mine tailings and then all the tailing soils were fully mixed (tailing properties: organic matter  $330 \pm 71.2$ mg kg<sup>-1</sup>, total P 50.3  $\pm$  10.6 mg kg<sup>-1</sup>, total N 120.7  $\pm$ 41.4 mg kg<sup>-1</sup>, Cu 31.29  $\pm$  3.81 mg kg<sup>-1</sup>, Cd 8.92  $\pm$  2.32 mg kg<sup>-1</sup>, Pb 924.57  $\pm$  61.23 mg kg<sup>-1</sup>, Zn 1,312.45  $\pm$ 52.44 mg  $kg^{-1}$ ). Natural soils was randomly collected from the Beigu Mountain in Linhai, Zhejiang, China and then all the natural soils were fully mixed (soil properties: organic matter 1,642.8  $\pm$  171.9 mg kg<sup>-1</sup>, total P 123.6  $\pm$ 21.7 mg kg<sup>-1</sup>, total N 734.3  $\pm$  162.8 mg kg<sup>-1</sup>). Another plant species survey was conducted at the Beigu, Yunfeng and Baiyunshan mountains on 28 October 2011. The eight species used in this study also occur in these mountain areas.

Seedlings of the eight species were transplanted into plastic containers (plots) placed outdoors (80  $\times$  80  $\times$  60 cm<sup>3</sup>) on 2–4 June 2009. Thirty-two plant seedlings were transplanted into each plot. We constructed 23 different communities to create four species richness levels (Table [1\)](#page-2-0): eight monocultures, five mixtures of two species, five mixtures of four species and one mixture of eight

<span id="page-2-0"></span>Table 1 The species combinations used at four different diversity levels BP Bidens pilosa Linn., PA Phytolacca americana Linn., CC Commelina communis Linn., MJ Mirabilis jalapa Linn., CA Chenopodium ambrosioides Linn., SN Solanum nigrum Linn., BC Brassica campestris Linn., XS Xanthium sibiricum Patrin ex Widder Diversity level  $1$  2 4 8 **BP** PA CC MJ CA SN BC XS  $BP + XS$  $PA + MJ$  $CC + CA$  $BP + PA$  $SN + BC$  $BP + CC + CA + SN$  $PA + MJ + CA + BC$  $XS + PA + SN + MJ$  $PA + BP + CC + BC$  $BP + CC + MJ + CA$  $BP + PA + CC + MJ$  $+ CA + SN + BC + XS$ 

species with five replicates. The species in the mixtures were planted at equal densities. Plots filled with natural soil were considered as the low stress treatment, and plots filled with tailing soil were considered as the high stress treatment. The alternative technique of a random selection of species from a total pool was not used because, with a limited series of communities, equal representation of the species at each species richness level was not guaranteed. Instead, combinations were chosen to guarantee every species could be selected at least once at each of four diversity levels. Plots were weeded weekly. In addition to  $H<sub>2</sub>SO<sub>4</sub>$  were added to 0.5 g of air-dried ground tailings in 10-ml digestion tubes. After heating, followed by the addition of 20 ml of distilled water and filtration, the contents of the tubes were transferred to 50-ml volumetric flasks. Total N in the filtrates was determined using the Berthelot reaction method (Page et al. [1982\)](#page-8-0).

The tolerance responses to stress of selected species were assessed by comparing the species biomass in low stress plots to that in high stress plots, as described previously by Suding et al. [\(2003](#page-8-0)). The biomass ratio was expressed as tolerance index (TI):

Average biomass of species in certain diversity treatment of high stress plots

Average biomass of species in the same diversity treatment of low stress plots

natural rainfall, water was added by artificially spraying during dry periods.

Plots were harvested about 12 months after the plot construction on 10 May 2010. All the aboveground and belowground biomass (living plants) were sorted to species, dried and weighed separately. In each plot, the species with the greatest biomass was identified as the dominant species of the plot.

Before the biomass harvesting, a water treatment experiment was conducted. The water content of each plot was determined 5 days after it received the same amount of water through artificial spraying (no water was added during the final 5 days before harvesting). Five soil cores to a depth of 0–20 cm were randomly collected in each plot. Five replicated samples from each plot were combined into one sample. Water content of the soil was calculated by comparing the weight of undried and dried soil [(weight of undried soil  $-$  weight of dried soil)/weight of dried soil]. Part of each sample was air-dried at room temperature, and then was homogenized and sieved  $(<2$  mm) to remove plant root and small stones to determine total N. Briefly, 1.0 g of  $K_2SO_4$  catalyst mixture and 5 ml of concentrated

We used the additive partitioning method (Loreau and Hector [2001](#page-8-0)) to quantify the selection and complementarity effects. The complementarity effect for a specific number of species N was N  $\triangle$ RYM, where  $\triangle$ RY was the average change in relative yield for all species in the mixture and  $\overline{M}$  was the average monoculture yield. The selection effect  $N \text{ cov}(\Delta RY)$ , M) was calculated as the covariance between the monoculture yield of species  $(M)$  and their change in relative yield in the mixture  $(\Delta RY)$  multiplied by N of the mixture.

The deviation  $(D)$  of the biomass of a species in a mixture from the biomass could be expected based on its monoculture biomass (Spaèková and Lepš [2001](#page-8-0)).

$$
D_{\text{mixture-monoculture}} = \frac{O_i - E_i}{E_i}
$$

where  $O_i$  is the observed biomass of species i in the mixture and  $E_i$  is the expected biomass, i.e. simply the monoculture biomass multiplied by the initial proportion of the species in the mixture. If  $D_{\text{mixture-monoculture}} > 0$ , the species expressed a better performance than the expected yield; if  $D_{\text{mixture-monoculture}} < 0$ , the species expressed a worse performance than the expected yield.

For each plot mixture, we calculated the over-yielding index (OI) as follows (Hector et al. [2002\)](#page-7-0):

$$
OI = Y/\text{MAX}(M_i)
$$

where Y was the biomass of a mixture;  $M_i$  was the biomass of ith species grown in the monoculture; OI was the ratio of the biomass of a mixture to the biomass of the most productive species in the mixture. For statistical analyses, log(OI) was used: if the biomass of each mixture reached the biomass of its most productive species in the monoculture, then the expectation of Log(OI) was zero.

ANOVA was used to test the effects of species richness and the identity of dominant species in mixtures in a sequential fitting order based on general linear model approaches through SPSS software (based on type I sum of squares; SPSS 11.5; SPSS, Chicago, Ill.). In the present study, simple regression analysis was used to examine the dependence of biomass on the selection or the complementarity effect, and also to examine the dependence of the OI and the TI on species richness.

#### Results

We explored the dependence of plot biomass on the selection and complementarity effects in low and high stress plots by using linear regression analysis (Fig. 1). The reverse biomass–complementarity effect relationships occur between low and high stress plots. The biomass negatively correlates with the complementarity effect in low stress plots but has a positive correlation in high stress plots. In contrast, the biomass positively correlates with the selection effect in low stress plots but negatively correlates in high stress plots. The results indicate that the main mechanism affecting biomass production shifts from the selection to the complementarity effect in response to increasing environmental stress.

Whether dominant species or species richness determine the plot biomass was also explored. In low stress plots, the biomass of mixtures are affected by the identity of domi-nant species (Table [2](#page-4-0);  $P < 0.01$ ) and positively correlate with the monoculture biomass of dominant species  $(r = 0.919, P < 0.0001, n = 15)$ . However, species richness does not affect the biomass of mixtures ( $P = 0.069$ ). In high stress plots, the biomass of mixtures positively correlates with species richness ( $r = 0.909$ ,  $P < 0.0001$ ,  $n = 15$ ) but has no relationship with dominant species  $(P = 0.082)$ . These results indicate that the main factor affecting the biomass of mixtures shifts from dominant species to species richness in response to increasing environmental stress. Moreover, both the selection and complementarity effects are significantly affected by the



Fig. 1 Relationship between biomass and complementarity or selection effect in a, c low and b, d high stress plots. Results for linear regression ( $n = 15$ ) are shown

<span id="page-4-0"></span>Table 2 Results of ANOVA for diversity effects on plot biomass, selection effect and complementarity effect in low and high stress plots based on type I sum of squares: arrows indicate significant

increase ( $\uparrow$ ) or decrease ( $\downarrow$ ) of the particular parameter with increased species richness or the monoculture biomass of dominant species



Significant P-values ( $P \lt 0.05$ ) in bold

identity of dominant species in low stress plots ( $P < 0.01$ ), while they are significantly affected by species richness in high stress plots ( $P < 0.05$ ).

The covariance between the monoculture biomass of species and their relative yields in the mixtures exhibit different relationships between low and high stress plots (Fig. 2). In low stress plots, the biomass proportions of species in the mixtures and their monoculture biomass have a positive relationship. Moreover, the selection effect positively correlates with the monoculture biomass of dominant species (Table 2;  $r = 0.928$ ,  $P < 0.0001$ ,  $n = 15$ ). However, in high stress plots, the monoculture biomass of species is not a good indicator of competitive ability in the mixtures. The species with a low or intermediate monoculture biomass frequently have high proportions in the mixtures, and the selection effect and the monoculture biomass of dominant species have no significant correlations (Table 2; Fig. 2).

The change in species interaction was also examined by the deviation between the observed and expected biomass  $(D_{\text{mixture-monoculture}})$  in this study. In low stress plots, the dominant species show a competitive advantage  $(D_{\text{mixture}})$  $_{\rm monoculture}>0$  but most of the non-dominant species show a competitive disadvantage  $(D_{\text{mixture-monoculture}} < 0)$  (Fig. [3](#page-5-0)). This proves that a dominant competitive interaction exists in low stress plots. However, in high stress plots, both the dominant species and most of non-dominant species have better performances than expected  $(D_{\text{mixture-monoculture}} > 0)$ . Most of mixtures have more biomass than the monoculture biomass of the most productive species in the mixtures (Fig. [4;](#page-5-0)  $Log(OI) > 0$ ). Moreover,  $Log(OI)$  positively correlate with species richness ( $r = 0.830$ ,  $P < 0.0001$ ,  $n = 15$ ). Only positive interaction (facilitation) can explain mixture plots having more biomass than the monoculture biomass of the most productive species in the mixtures.



Fig. 2 Biomass proportions of species in mixtures as a function of monoculture biomass; values are mean  $\pm$  SE ( $n = 10$  for PA and BP;  $n = 9$  for *MJ*, *CA* and *SN*;  $n = 8$  for *CC* and *BC*;  $n = 7$  for *XS*) in a low and b high stress treatments. Abbreviations for species names are listed in Table [1](#page-2-0)

The facilitation can be crucial in high stress environments by improving environmental conditions. In high stress plots, the nitrogen and water retention capability of mine tailings increase with species richness, as total N and water content positively correlate with species richness (Table [3;](#page-6-0) total N,  $r = 0.847$ ,  $P < 0.0001$ ,  $n = 15$ ; water

<span id="page-5-0"></span>

Fig. 3 Deviation ( $D_{\text{mixture-monoculture}}$ ) of observed biomass of dominant or non-dominant species from their expected biomass in a low and b high stress mixture plots. Expected biomass was monoculture biomass multiplied by the initial proportion of the species in the mixture



Fig. 4 Dependence of log[over-yielding index $(OI)$ ] on the number of species in low and high stress mixture plots. Only the regression for high stress plots is significant ( $r = 0.83$ ,  $n = 15$ ,  $P < 0.0001$ )

content,  $r = 0.862$ ,  $P < 0.0001$ ,  $n = 15$ ). The facilitations between species may mitigate the impact of environmental stress as TI values of all eight species increase with species richness (Fig. [5\)](#page-6-0).

#### **Discussion**

In our controlled pot experiment, we find a positive dependence of biomass on the selection effect in low stress plots, while, in high stress plots, the biomass positively depends on the complementarity effect. The relationship between the monoculture biomass of species and their relative yields in the mixtures exhibits different patterns. In low stress plots, there is a positive covariance between them. For example, the most productive species in monocultures also dominated in mixtures. Conversely, in high stress plots, most of the unproductive species in monocultures perform better than expected in mixtures  $(D_{\text{mixture}})$  $_{\rm monoculture}>0$ ), and most of the mixtures have more biomass than the monocultures of their most productive species  $[log(OI) > 0]$ . However, these better performances are not found in low stress plots. These differences occur because species interactions varied with abiotic stress: the competition prevails in low stress plots, while the facilitation dominates in high stress plots.

The relationship between the mixture and monoculture yield has become a critical issue in interpreting biodiversity experiments (Špaèková and Lepš<sup>[2001](#page-8-0)</sup>; Hector et al. [2002](#page-7-0)). In our study, the relationship between the mixture and monoculture yield shifts with species interactions that varied with abiotic stress. In low stress plots, the competition is the principal species interaction and the dominant species determines plot biomass. A positive selection effect occurred as species had a positive covariance between the relative yield and the monoculture yield (Fargione et al. [2007](#page-7-0)). Furthermore, the monoculture biomass of species is a good indicator of competitive dominance in the mixtures. When the species that are abundant in monoculture dominate the mixtures, strong dominance leads to a high selection effect and plot biomass; when the species that are not abundant in monoculture dominate the mixtures, weak dominance leads to a low selection effect and plot biomass. Therefore, the plot biomass has a positive dependence on the selection effect. In high stress plots, the principal species interaction shifts to facilitation. Both dominant species and non-dominant species perform better than their expected performances ( $D_{\text{mixture-monoculture}} > 0$ ). A positive complementarity effect occurs as species, on average, have higher relative yield than the expected value (Fargione et al. [2007\)](#page-7-0). The increase in the TI and the OI indicate that the function of facilitation is strengthened with the increase of plot species richness. The higher the species richness, the more positive the deviations of the average relative yield from the expected value and the greater the plot biomass. Consequently, the plot biomass positively links with the complementarity effect. Based on our results, we suggest that the positive covariance between the relative yields of species in the mixtures and their monoculture

marcule significant increase ()) or are pararelate parameter while increased species fromcos of anomalism of dominant species					
Source of variation	df	Total N		Water content	
		F	P	F	P
Low stress plots					
Richness	$\overline{c}$	1.367	0.353	2.644	0.185
Dominant species	4	2.880	0.172	3.966	0.103
Richness $\times$ dominant species	4	1.448	0.364	1.114	0.459
High stress plots					
Richness	$\mathfrak{D}$	93.771	< 0.05	152.496	< 0.05
Dominant species	6	6.944	0.283	8.912	0.251
Richness $\times$ dominant species		12.739	0.209	11.071	0.224

<span id="page-6-0"></span>Table 3 Results of ANOVA for diversity effects on total N and water content in low and high stress plots based on type I sum of squares: arrows indicate significant increase (1) of the particular parameter with increased species richness or the monoculture biomass of dominant species

Significant P values ( $P < 0.05$ ) in bold



Fig. 5 Dependence of the tolerance index on the number of species in a monoculture or mixture for eight species plots used in experiments. Regression lines are shown for all the eight species, although not all are significant (see text for details). Species abbreviations are listed in Table [1](#page-2-0)

yields should disappear when species interactions shift from competition to facilitation.

Our results are consistent with the stress gradient hypothesis, i.e. that competition and facilitation vary inversely across gradients of abiotic stress. Facilitation is the dominant interaction under highly stressful conditions (Bertness and Callaway [1994](#page-7-0)), while competition dominates in low stress environments (Bertness and Callaway [1994;](#page-7-0) Brooker and Callaghan [1998](#page-7-0); Callaway et al. [2002](#page-7-0); Chu et al. [2008](#page-7-0)). However, species interactions may vary from competition to facilitation as a function of the tolerance of a species to particular abiotic conditions (Bertness et al. [1992](#page-7-0); Hacker and Bertness [1999](#page-7-0); Corcket et al. [2003](#page-7-0); Lortie et al. [2004\)](#page-8-0). In this study, the monoculture biomasses of eight species were greatly restrained and decreased greatly (the range of the decrease ratios of eight species is 54.2–83.8 %) in high stress plots, which indicate that the abiotic stress of mine tailings was close to their maximum. At the Huangyan Pb/Zn mine tailings site, bare mine tailings soils have more heavy metals and less nutrients than those soils that have been phytoremediated for many years (Wang et al. [2011\)](#page-8-0). The eight species used in this study have grown on phytoremediated and surrounding areas for many years. The environment of bare mine tailing soil should be very harsh for these species that have adapted to the soils of phytoremediated and surrounding areas. Consequently, facilitation is the main species interaction in high stress plots. However, we speculate that if highly stress-tolerant species are experimental targets where bare mine tailing soil is not very harsh for them, then competition may be expected in high stress plots (Liancourt et al. [2005](#page-8-0)). Pennings et al. [\(2003](#page-8-0)) found that competition remains the main interaction of salt-tolerant species under the great stress of soil salinity. Therefore, the two cases observed in this study may be the end points of a continuum from ''optimum environment'' to "extreme environment" for the eight plant species. Consequently, the shift between the selection and complementarity effects depends on experimental abiotic stress relative to the optimum conditions of the target species.

In our study, the diversity–biomass relationships exhibit different patterns between low and high stress plots. In low stress plots, the selection effect results in no significant relationship between the plot biomass and species richness. The plot biomass positively correlates with the selection effect. As a result, the plot biomass is determined by the characteristic of the dominant species and has no close link with species richness. For example, some species that are abundant in monocultures dominate some of the low diversity plots, which leads to low diversity plots with a high selection effect and high biomass, while other species that are not abundant in monocultures dominate some of the high diversity plots, which leads to high diversity plots with a low selection effect and low biomass.

In contrast, in high stress plots, the plot biomass and species richness have a significantly positive relationship.

<span id="page-7-0"></span>The complementarity effect is the main mechanism which leads to a positive relationship between the plot biomass and species richness. Higher diversity plots have more facilitation that leads to a high complementarity effect and high plot biomass. Consequently, a positive relationship occurs in high stress plots. The facilitation may have been driven by the water retention of the plant communities, as we found that soil moisture increases with species richness. Moreover, the nutrient pool is significantly improved with the increase of species richness. The improvement may be the result of exudates released by the roots and microbial metabolites (Marschner and Romheld [1983;](#page-8-0) Leyval and Berthelin [1993\)](#page-8-0), as the underground biomass of plots increases with species richness (data not shown). Many previous biodiversity experiments in natural environments did exhibit different diversity–biomass relationships (Loreau and Hector [2001;](#page-8-0) Fridley 2002; van Ruijven and Berendse [2003;](#page-8-0) Hooper and Dukes 2004; Spehn et al. [2005](#page-8-0)), while simple positive relationships generally occurred in harsh environments (Mulder et al. [2001](#page-8-0); Callaway et al. 2002; Wang et al. [2011](#page-8-0)). Based on these biodiversity experiments and the results from this study, we speculate that the complex diversity–biomass relationships should shift to simple positive relationships when the principal mechanism shifts from the selection effect to the complementarity effect.

There are two important points to note from our results. First, we report results based on a 1-year experiment. But both the selection and complementarity effects are likely to change over time. In a 10-year biodiversity experiment, Fargione et al. (2007) found that the selection effect decreased and the complementarity effect increased over time. Second, the pot experiment limits environmental fluctuations and spatial heterogeneity, which are important for resource partitioning between species under natural conditions (Špaèková and Lepš $2001$ ). As a result, niche differentiation, an important component of the complementarity effect, may be restrained in our pot experiment.

In conclusion, our study finds a shift in the mechanism of the biodiversity effect on productivity along an abiotic stress gradient. The shift occurs because species interaction varied with the magnitudes of abiotic stress. Competition prevails in low stress plots, which leads to the selection effect positively determining the plot biomass. Facilitation becomes the principal species interaction in high stress plots, and the biomass of plots positively depends on the complementarity effect. Although the pot experiment has its limitation for resource partitioning, it indicates that the shift in species interactions caused by the change in abiotic stress will determine the principal mechanism of the biodiversity effect.

Acknowledgments Thanks to C. B. Zhang for help with harvesting, and Y. Ge provided helpful comments on the manuscript. This study

was completed within the experimental center of Taizhou University. These experiments comply with the current laws of the China. This research was financially supported by the National Natural Science Foundation of China (project 31000256) and by the Natural Science Foundation of Zhejiang Province, China (Y5100016).

### References

- Bertness MD, Callaway RM (1994) Positive interactions in communities. Trends Ecol Evol 9:191–193
- Bertness MD, Gough L, Shumway SW (1992) Salt tolerances and the distribution of fugitive salt marsh plants. Ecology 73:1842–1851
- Brooker RW, Callaghan TV (1998) The balance between positive and negative plant interactions and its relationship to environmental gradients: a model. Oikos 81:196–207
- Callaway RM (2007) Positive interactions and interdependence in plant communities. Springer, Dordrecht
- Callaway RM, Brooker RW, Choler P, Kikvidze Z, Lortie CJ, Michalet R (2002) Positive interactions among alpine plants increase with stress. Nature 417:844–848
- Chu CJ, Maestre FT, Xiao S, Weiner J, Wang YS, Duan ZH, Wang G (2008) Balance between facilitation and resource competition determines biomass–density relationships in plant populations. Ecol Lett 11:1189–1197
- Corcket E, Liancourt P, Callaway RM, Michalet R (2003) The relative importance of competition for two dominant grass species, as affected by environmental manipulations in the field. Ecoscience 10:186–194
- Dukes JS (2001) Productivity and complementarity in grassland microcosms of varying diversity. Oikos 94:468–480
- Engelhardt KAM, Ritchie ME (2001) Effects of macrophyte species richness on wetland ecosystem functioning and services. Nature 411:687–689
- Fargione J, Tilman D, Dybzinski R, Lambers JHR, Clark C, Harpole WS, Knops JMH, Reich PB, Loreau M (2007) From selection to complementarity: shifts in the causes of biodiversity–productivity relationships in a long-term biodiversity experiment. Proc R Soc B 274:871–876
- Fridley JD (2001) The influence of species diversity on ecosystem productivity: how, where, and why? Oikos 93:514–526
- Fridley JD (2002) Resource availability dominates and alters the relationship between biodiversity and ecosystem productivity in experimental plant communities. Oecologia 132:271–277
- Hacker SD, Bertness MD (1999) Experimental evidence for factors maintaining plant species diversity in a New England salt marsh. Ecology 80:2064–2073
- He ZL, Yanga XE, Stoffellab PJ (2005) Trace elements in agroecosystems and impacts on the environment. Rev J Trace Elem Med Biol 19:125–140
- Hector A (2002) Biodiversity and the functioning of grassland ecosystems: multi-site comparisons. In: Kinzig AP, Pacala S, Tillman D (eds) The functional consequences of biodiversity. Princeton University Press, Princeton, N.J., pp 71–95
- Hector A, Bazeley-White E, Loreau M, Otway S, Schmid B (2002) Overyielding in grassland communities: testing the sampling effect hypothesis with replicated biodiversity experiments. Ecol Lett 5:502–511
- Hooper DU, Dukes JS (2004) Overyielding among plant functional groups in a long-term experiment. Ecol Lett 7:95–105
- Hooper DU, Vitousek PM (1998) Effects of plant composition and diversity on nutrient cycling. Ecol Monogr 68:121–149
- Hooper DU, Chapin FS III, Ewel JJ, Hector A, Inchausti P, Lavorel S (2005) Effects of biodiversity on ecosystem functioning: a consensus of current knowledge. Ecol Monogr 75:3–35
- <span id="page-8-0"></span>Huston MA (1997) Hidden treatments in ecological experiments: reevaluating the ecosystem function of biodiversity. Oecologia 108:449–460
- Huston MA, McBride AC (2002) Evaluating the relative strengths of biotic versus abiotic controls on ecosystem processes. In: Loreau M, Naeem S, Inchausti P (eds) Biodiversity and ecosystem functioning: synthesis and perspectives. Oxford University Press, New York, pp 47–60
- Huston MA, Aarssen LW, Austin MP (2000) No consistent effect of plant diversity on productivity. Science 289:1255a
- Kaiser J (2000) Rift over biodiversity divides ecologists. Science 289:1282–1283
- Kikvidze Z, Pugnaire FI, Brooker RW, Choler P, Lortie CJ, Michalet R, Callaway RM (2005) Linking patterns and processes in alpine plant communities: a global study. Ecology 86:1395–1400
- Leyval C, Berthelin J (1993) Rhizodeposition and net release of soluble organic compounds by pine and beech seedlings inoculated with rhizobacteria and ectomycorrhizal fungi. Biol Fertil Soils 15:259–267
- Liancourt P, Callaway RM, Michalet R (2005) Stress tolerance and competitive-response ability determine the outcome of biotic interactions. Ecology 86:1611–1618
- Loreau M (1998) Biodiversity and ecosystem functioning: a mechanistic model. Proc Natl Acad Sci USA 95:5632–5636
- Loreau M (2000) Biodiversity and ecosystem functioning: recent theoretical advances. Oikos 91:3–17
- Loreau M, Hector A (2001) Partitioning selection and complementarity in biodiversity experiments. Nature 412:72–76
- Lortie CJ, Brooker RW, Kikvidze Z, Callaway RM (2004) The value of stress and limitation in an imperfect world: a reply to Körner. J Veg Sci 15:577–580
- Maestre FT, Valladares F, Reynolds JF (2005) Is the change of plant– plant interactions with abiotic stress predictable? A metaanalysis of field results in arid environments. J Ecol 93:748–757
- Marschner H, Romheld V (1983) In vivo measurement of rootinduced pH changes at the soil–root interface: effect of plant species and nitrogen source. Plant Physiol 111:241–251
- Mulder CPH, Uliassi DD, Doak DF (2001) Physical stress and diversity–productivity relationships: the role of positive interactions. PNAS 98:6704–6708

Naeem S (2000) Reply to Wardle et al. Bull Ecol Soc Am 81:241–246

- Page AL, Miller RH, Keeney DR (1982) Methods of soil analysis chemical and microbiological properties. ASA and SSSA, Madison, WI
- Pennings SCER, Seling LTH, Bertness MD (2003) Geographic variation in positive and negative interactions among salt marsh plants. Ecology 84:1527–1538
- Pfisterer AB, Joshi J, Schmid B, Fischer M (2004) Rapid decay of diversity–productivity relationships after invasion in experimental plant communities. Basic Appl Ecol 5:5–14
- Schwartz MW, Brigham CA, Hoeksema JD (2000) Linking biodiversity to ecosystem function: implications for conservation ecology. Oecologia 122:297–305
- $\text{\r{S}}$ paèková I, Lepš J (2001) Procedure for separating the selection effect from other effects in diversity–productivity relationship. Ecol Lett 4:585–594
- Spehn EM, Hector A, Joshi J, Scherer-Lorenzen M, Schmid B, Bazeley-White E (2005) Ecosystem effects of biodiversity manipulations in European grasslands. Ecol Monogr 75:37–63
- Suding KN, Goldberg D, Hartman KM (2003) Relationships among species traits: separating levels of response and identifying linkages to abundance. Ecology 84:1–16
- Tilman D (1988) Plant strategies and the dynamics and structure of plant communities. Princeton University Press, Princeton, N.J.
- van Ruijven J, Berendse F (2003) Positive effects of plant species diversity on productivity in the absence of legumes. Ecol Lett 6:170–175
- Vega FA, Covelo EF, Andrade ML (2006) Competitive sorption and desorption of heavy metals in mine soils: influence of mine soil characteristics. J Colloid Interface Sci 298:582–592
- Wang J, Zhang CB, Ke SS, Qian BY (2011) Different spontaneous plant communities in Sanmen Pb/Zn mine tailing and their effects on mine tailing physico-chemical properties. Environ Earth Sci 62:779–786
- Wardle DA (1999) Is "sampling effect" a problem for experiments investigating biodiversity–ecosystem function relationships? Oikos 87:403–407
- Wardle DA, Huston MA, Grime JP, Berendse F, Garnier E, SetaÈlaÈ H (2000) Biodiversity and ecosystem functioning: an issue in ecology. Bull Ecol Soc Am 81:235–239