# The Properties and Mechanism of CuO Modified Carbon Nanotube for $NO_x$ Removal

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Abstract Carbon nanotube (CNT) supported copper oxide catalysts were prepared, and the morphology and structure of the catalyst were characterized by using scanning electron microscopy, transmission electron microscopy, X-ray diffraction and temperature-programmed desorption experiments. The CuO/CNT catalysts demonstrated high catalytic activities during the selective catalytic reduction (SCR) of NO with NH<sub>3</sub> over a temperature range of 150-250 °C. The amount of NH<sub>3</sub> adsorbed on the catalyst surface was greater than that of NO or NO +  $O_2$  adsorbed on the catalyst surface. These results suggest that the SCR reaction might proceed on the surface of the CuO/CNT catalysts and occur between the adsorbed ammonia and gas phase NO or weakly adsorbed NO. The CuO/CNT catalysts exhibited good stability at low temperatures, which makes them suitable for potential applications in industry.

**Keywords** Carbon nanotube · Copper oxide · NO reduction · Ammonia

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### **1** Introduction

Atmospheric pollutants (sulfur dioxide, nitrogen oxides  $(NO_x)$ ) are generated mainly from the burning of coal and oil. In particular,  $NO_x$  harm to the environmental because they generate acid rain, photochemical smog and cause greenhouse effects as well as destroy the ozone layer [1, 2]. Many methods have been developed to reduce the harmful effects of NO<sub>x</sub>. The SCR of NO<sub>x</sub> with ammonia might be the most effective methods for the reduction of NO<sub>x</sub>. Developing a suitable catalyst for SCR reaction remains a key problem. Previous studies indicated that titanium dioxide loaded with vanadium catalysts show high efficiency for reduction nitrogen oxides above 350 °C [3, 4]. However, this catalyst has many disadvantages, including high operational temperatures and the toxicity of vanadium species [5]. Therefore, many new catalysts have been developed to avoid these shortcomings [6-8]. Many reported catalysts, such as MnO<sub>2</sub>/ Al<sub>2</sub>O<sub>3</sub>, CuO/activated carbon (AC) and Fe<sub>2</sub>O<sub>3</sub>/AC, demonstrate high activity in SCR reaction at 120–250 °C [9–12].

Recently, copper oxide catalysts have attracted considerable attention because they exhibit high low-temperature activity for catalytic removal of NO [13–18], especially activated carbon as a support. However, we could not recognize the catalytic nature for this catalyst due to the complexity of surface chemistry and pore structure of activated carbon. Carbon nanotube has pure surface and simple structure, and it is useful for us to know the catalytic nature for SCR. The CNT-supported metal oxides such as  $CeO_2$ ,  $MnO_X$ , etc. have been reported for low temperature deNO<sub>x</sub> catalysts and showed good catalytic activity [19–23]. In view of peculiar character of carbon surface on carbon nanotube for removal NO [8, 24], it is important theory for development low-temperature SCR catalysts to modify carbon nanotubes with CuO. In the present work,

the properties and mechanism of CNT-supported CuO catalysts (CuO/CNT) were studied for the low-temperature SCR of NO with NH<sub>3</sub>, which contributes to understanding the catalytic nature of carbon-based catalysts.

#### 2 Experimental

#### 2.1 The Catalyst Preparation

The raw CNT (Tsinghua University) samples were purified by the same method as ref. mentioned [25]. The CuO/CNT catalysts were prepared by pore volume impregnation of the purified CNT with an aqueous solution of cupric nitrate solution. The catalysts were dried at 60 °C overnight and then at 110 °C for 5 h, followed by calcined in argon stream at 250 °C for 2 h. For comparison, the Cuo/TiO<sub>2</sub> (ISK, Japan) catalysts were obtained by the same method.

#### 2.2 Activity Tests

The SCR activity tests were carried out in a fixed-bed glass reactor (8 mm in inner diameter and 600 mm in length). The flue gas was simulated by NO in Ar, pure  $O_2$  and pure Ar, SO<sub>2</sub> (When used) in Ar and NH<sub>3</sub> in Ar was used as reductive gas. All the gases were controlled by mass flow controllers. Concentration of NO, N<sub>2</sub>O, NO<sub>2</sub>, SO<sub>2</sub> and O<sub>2</sub> at the inlet and the outlet of the reactor were simultaneously monitored by an online combustion gas analyzer (Testo-pro 350, Germany) equipped with NO, NO<sub>2</sub>, SO<sub>2</sub>, O<sub>2</sub> and analyzer (PGD-100, Austria) equipped with N<sub>2</sub>O.

#### 2.3 The Catalyst Characterization

The CuO/CNT catalysts were characterized by X-ray diffraction(XRD) patterns, scanning electron microscopy (SEM, JEOL JSM-6700F), and transmission electron microscopy (TEM, JEOL JEM 3010) equipped with energy dispersive X-ray spectroscopy(EDS).

The temperature-programmed desorption of NH<sub>3</sub> (NH<sub>3</sub>-TPD) and NO or NO +  $O_2$  was performed in a fixed-bed quartz reactor. A typical experiment used at a 200 mg sample and a gas flow rate of 100 ml/min. The catalyst was pre-treated under Ar at 300 °C for 2 h before being subjected to the adsorption of 1,000 ppm of a sample gas for 1 h at 30 °C, and TPD was measured at 10 °C/min under Ar from 30 to 960 °C.

#### **3** Results and Discussion

Figure 1 illustrates the SEM, TEM and XRD patterns of the 10 wt% CuO/CNT catalyst. It is obvious that the

diffraction peaks  $(2\theta = 26.0, 42.8)$  can be indexed to CNT [20], and the diffraction peaks  $(2\theta = 36.3)$  were attributed to hexagonal phase of CuO (JCPDS Card No. 45-0937) [26]. The average size of CuO particles on the catalysts surface was approximately 6–10 nm, which was obtained by analysis of the appropriate XRD date and agreed with the TEM results.

Figure 2a summarizes the changes in the SCR activity at different CuO loading relative to the weight of CuO/CNT catalyst at 200 °C. The catalytic activity of CuO/CNT catalyst increased when the CuO loading was increased within the range of 1-10 wt%. NO conversion reached 88.5 % when the CuO loading was 10 wt%. However, NO conversion decreased when CuO loading was greater than 10 wt%. The decrease in the catalytic activity might be due to aggregation of the copper oxide particles on the CNT surface. N<sub>2</sub>O has not been observed in this reaction process and only a small amount of NO<sub>2</sub> (less than 10 ppm) was observed. To further exhibit the excellent catalytic properties of CuO/CNT catalysts, SCR activity of a commercial titanium dioxide-supported CuO catalysts (CuO/TiO<sub>2</sub>) was investigated and shown in Fig. 2b. It is obvious that the SCR activity of the CuO/CNT catalytst is much higher than that of the CuO/TiO<sub>2</sub> catalyst.

Figure 3 reveals the performance of the 10 wt% CuO/ CNT catalyst during the reduction of NO<sub>x</sub> with a reaction temperature between 120 and 250 °C. These results demonstrated that the NO conversion over the 10 wt% CuO/ CNTs catalyst increased when the reaction temperature was increased. NO conversion reached approximately 88.5 % when the reaction temperature was increased to 200 °C. However, when the reaction temperature was increased to 250 °C, NO conversion declined to approximately 70.5 % because the oxidation of ammonia was more favorable at increased reaction temperatures which led to decreased reduction of NO by ammonia. The appearance of a small amount of  $N_2O$  and  $NO_2$  in the reaction mixture confirmed this process.

The effect of the flue gas space-velocity on NO conversion with 10 wt% CuO/CNT catalyst under the reaction temperature of 200 °C is shown in Fig. 4. The NO conversion reached 97.3 % when the flue gas space velocity was  $10,000 \text{ h}^{-1}$ . The NO conversion also exhibited a negative trend when the flue gas space velocity was increased. Using a small space velocity can increase the amount of the flue gas adsorbed on the surface and provides a long interval for the flue gas to react with an active component. However, NO conversions can high reached 88.5 % when the space velocity was  $30,000 \text{ h}^{-1}$ .

From a theoretical and technical point of view, the stability of the CuO/CNT catalysts must be studied. Figure 5 shows the stability of the 10 wt% CuO/CNT catalyst at 200 °C. The CuO/CNT catalyst demonstrates excellent



Fig. 1 SEM (a, b) and TEM (c) of the 10 wt% CuO/CNT catalysts and X-ray diffraction of copper oxide catalysts support on carbon nanotubes (d)



**Fig. 2** a Effect of CuO loading on on the activity of CuO/CNTs catalysts at 200 °C; **b** NO conversion of CuO/CNT catalysts and CuO/TiO<sub>2</sub> at 200 °C. Reaction conditions: 450 ppm NO, 500 ppm NH<sub>3</sub>, 5 vol% O<sub>2</sub>, WHSV, 30,000 h<sup>-1</sup>, Catalyst weight: 200 mg

catalytic activity and NO conversion efficiency keeps 88.5 % over a 72 h of use. These results indicate that the CuO/CNT catalysts have a moderately good stability.

To understand the influence of  $SO_2$  on the CuO/CNT catalyst,  $SO_2$  transient response experiments consisting of five consecutive parts (Fig. 6) were carried out at 200 °C

using the 10 wt% CuO/CNT catalyst. The concentrations of NO and SO<sub>2</sub> were monitored at the outlet during the experiments. In part I, NO conversion was approximately 88.5 % in the absence of SO<sub>2</sub>; however, this value decreased to 9 % after introducing SO<sub>2</sub> into the system in part II, and further increases in the SO<sub>2</sub> concentration



Fig. 3 Effect of reaction temperature on the activity of the 10 wt% CuO/CNTs catalysts. Reaction conditions: 450 ppm NO, 500 ppm NH<sub>3</sub>, 5 vol%  $O_2$ , WHSV, 30,000 h<sup>-1</sup>, Catalyst weight: 200 mg



Fig. 4 Effect of space velocity on the activity of the 10 wt% CuO/ CNTs catalysts. Reaction conditions: 450 ppm NO, 500 ppm NH<sub>3</sub>, 5 vol%  $O_2$ , Reaction temperature: 200 °C

exhibit similar results. When SO<sub>2</sub> was removed from the feed, there was a small increase in the NO conversion (as shown in part III). After the catalyst was heated in Ar from 200 to 450 °C, which was followed by cooling to the original temperature of 200 °C (part IV). Subsequently, when the feed gas was re-subjected to the conditions in part I, the NO conversion returned to a consistent value of 40 % (part V), which was much lower than the conversion in part I. This conclusion agrees with previous observations with CuO/CNT catalysts at low temperature [11]. The influence of H<sub>2</sub>O on the SCR activity over CuO/CNT catalyst was investigated and is shown in Fig. 6b. In the absence of H<sub>2</sub>O, the NO conversion over the 10 wt% CuO/CNT catalyst was 88.5 %, and a slight decrease after introducing



Fig. 5 A test of the stability of the 10 wt% CuO/CNTs catalysts for SCR reaction. Reaction conditions: 450 ppm NO, 500 ppm NH<sub>3</sub>, 5 vol%  $O_2$ , WHSV, 30,000 h<sup>-1</sup>, Catalyst weight: 200 mg, Reaction temperature: 200 °C



**Fig. 6** Effect of SO<sub>2</sub> (**a**) and H<sub>2</sub>O (**b**) on the activity of the 10 wt% CuO/CNTs catalysts. Reaction conditions: 450 ppm NO, 500 ppm NH<sub>3</sub>, 5 vol% O<sub>2</sub>, 400 ppm SO<sub>2</sub> (when used), 7.0 vol% H<sub>2</sub>O (when used), WHSV, 30,000 h<sup>-1</sup>, Catalyst weight: 200 mg, reaction temperature, 200 °C



Fig. 7 NH<sub>3</sub>- TPD profiles of the 10 wt% CuO/CNT catalyst



Fig. 8 NO + O<sub>2</sub>- TPD profiles of the CNT (a) and 10 wt% CuO/ CNT catalyst (b)

7 vol%  $H_2O$ . This result shows that  $H_2O$  did not show obvious inhibition on the SCR activity to this SCR catalyst [19], and the inhibition of  $H_2O$  may be due to the competitive adsorption between  $H_2O$  and  $NH_3$  on the active sites of the catalyst surface. These observations agree well with previous reports [21].

To observe the adsorption state of  $NH_3$  on the surface of catalyst, the TPD curves were recorded after  $NH_3$  was absorbed into the 10 wt% CuO/CNT catalysts. Figure 7 exhibits two distinct peaks, which are centered at approximately 100 and 282 °C and a very strong peak centered at 710 °C, which suggests that there are at least two  $NH_3$  species adsorbed on the catalyst. The low-temperature (LT) peak was associated with the ammonium ions adsorbed on BrΦnsted acid sites, and the high-temperature (HT) peak might be attributed to the molecularly ammonia from Lewis acid sites [8, 25]. The HT peaks are much larger than the LH peak, which suggest that the molecular  $NH_3$  was major form of  $NH_3$  adsorbed on the CuO/CNT catalysts is.



Fig. 9 NO TPD profiles of the CNT (a) and 10 wt% CuO/CNT catalyst (b)

NO and NO +  $O_2$  TPD curves over CNT and CuO/CNT catalysts are summarized in Figs. 8 and 9. The amount of desorbed NO on the CNT surface was small in the presence and in the absence of  $O_2$ . The amount of NO desorbed over the CuO/CNT catalysts was larger when  $O_2$  was present. However, the amount of NH<sub>3</sub> desorped over the CuO/CNT catalysts was larger than the amount of desorbed NO in the presence or absence of  $O_2$ . These results suggest that the SCR reaction proceeds on the CuO/CNT catalysts between the adsorbed ammonia and NO from the gas phase or weakly adsorbed NO.

#### 4 Conclusions

Copper oxide supported on carbon nanotube developed for the SCR of NO with ammonia demonstrates excellent catalytic activity over the temperature range of 150–250 °C. The catalysts exhibit good stability. The SCR reaction proceeds between adsorbed ammonia and NO that exists in the gas phase or weakly adsorbed NO. The best NO conversion of 88.5 % was obtained with the 10 wt% CuO/CNT at 200 °C and a GHSV of 30,000 h<sup>-1</sup>.

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